

Structure and Spatial Pattern of Vegetation in a Fragment of Atlantic Forest

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Abstract

The degradation of the Atlantic Forest is a reflection of the occupation and irrational exploitation of natural resources, which resulted in a drastic reduction of vegetation cover and fragmentation of these habitats. Thus, the objective of this study was to evaluate the species diversity, horizontal structure and spatial pattern of a fragment of Atlantic. Ten plots with dimensions of 10 m x 25 m (250 m²), were distributed in a completely random manner in the area. The data were obtained from the identification and measurement of the arboreal individuals with circumference at the breast height (Cbh) at 1.30 m of the soil \geq 15 cm. The total height was estimated by using a 6 m long ruler positioned near the individual to be measured. The horizontal structure was calculated based on the following phytosociological parameters: density, frequency, dominance, importance value and diversity indices (Shannon Index - H' and Pielou Equability - J). The spatial distribution pattern of the species was analyzed for the ten of the highest absolute density observed in the phytosociological analysis of the fragment. For this, we used the MacGuinnes index. The diversity Shannon Index was 3.37 nats. ind.⁻¹ and the Pielou equability index of 0.83. In general, the species presented an aggregate pattern of distribution in the area, with emphasis on the species *Miconia prasina* (Sw.) DC. and *Myrcia splendens* (Sw.) DC. and *Byrsonima crassifolia* that presented a random pattern, this information is necessary for future management and conservation actions in the region.

Key words: floristic, distribution of species; tropical forest; phytosociology.

INTRODUCTION

Native forests are still the main targets for environmental degradation in the world, and those that are classified as tropical suffer from even more intense and severe anthropogenic disturbances (Vaidyanathan et al., 2010, Estes et al., 2011; Morris, 2010). As a result, they tend to go through complex and gradual successional stages that aim to reestablish the various ecosystem processes previously lost and that maintain the self-sustaining functioning of this environment (Feeley et al., 2011).

In the Atlantic Forest, due to the marked anthropogenic action, it is challenging to maintain and conserve the biodiversity of the Atlantic Forest, which currently reflects the consequences of its degradation initiated since the colonial period (Turchetto et al., 2017). This phytogeographic domain presents a very diverse fauna and flora and was pointed out as the most devastated in Brazil according to data from the IBGE (2018). Carvalho et al. (2016) emphasize that the process of forest fragmentation causes physical and biological changes in the environment.

The Atlantic Forest is considered by the scientific community as an area of importance for the conservation of global biodiversity (Alves-Costa et al., 2008), because it has a high diversity and is affected by intense anthropic pressure. Therefore, studies that attenuate the effects of degradation in this phytogeographic domain are of great relevance.

Thus, Phytosociology becomes a relevant tool in obtaining information of great importance, since it helps in procedures that aim at the recovery of degraded environments, the production of seedlings, the recognition of endangered species, besides influencing decision-making on the most appropriate management actions, that should be done in a certain area for conservation purposes (Felfili and Venturoli, 2000; Brito et al., 2007).

When one wishes to infer about the structure and behavior of forest communities and to portray the relationship between vegetation, soil and climate, the indexes titled phytosociological parameters are used, which allow predicting characteristics such as the stage of development of the community but also on the quality and productivity of the analyzed environment (Souza and Soares, 2013; Chaves *et al.*, 2013).

Degradation processes resulting from natural events or human activities entail the fragmentation of these environments (Gris *et al.*, 2014; Zanini *et al.*, 2014), reflecting directly on the landscape structure, as well as results in different levels and degrees of disturbance, implying variation in the frequency and abundance of pioneer species and vegetation patterns (Lasky *et al.*, 2013). In view of the above, the objective of this work was to analyze the species diversity, the horizontal structure and the spatial pattern of a fragment of the Atlantic Forest located in Dois Irmãos State Park, Recife, Pernambuco, Brazil.

2 MATERIAL AND METHODS

Study Area

The research was developed in a fragment of Atlantic Forest of 774.09 ha, located in Dois Irmãos State Park (PEDI), Recife, Pernambuco, Brazil. The Park is a Conservation Unit of Integral Protection and has a total area of 1.158 ha, composed of two fragments in different successional stages, being the Mata de Dois Irmãos with 384.42 ha of mature forest and Brejo dos Macacos Farm with 774.09 ha of secondary forest whose area was incorporated into the PEDI by Decree 40.547 in 2014, being the area under study (Figure 1).

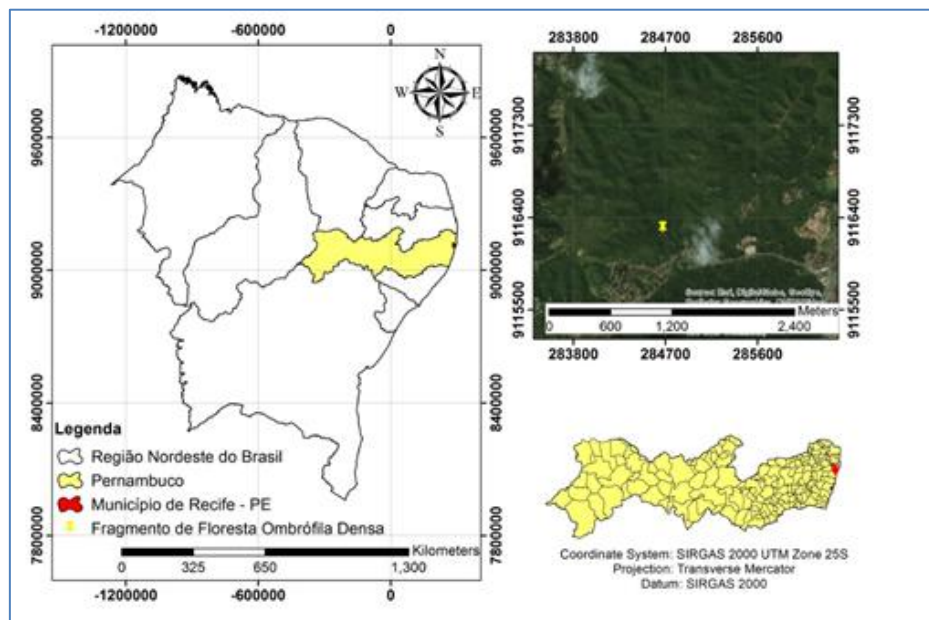


Fig. 1: Geographic location of the fragment of Ombrophylous Dense Forest, Recife, Pernambuco, Brazil.

The climate of the region is of type As' (tropical rainy) according to the classification of Köppen, with high temperatures throughout the year and rainy season concentrated between the months of March and August (Alvares *et al.*, 2013).

Data Collection

The sampling process consisted of a floristic and parametric survey of the vegetation, where 10 plots with dimensions of 10 m x 25 m (250 m²) were distributed randomly in the fragment, approximately 20 m from the edge of the remnant. All tree individuals with Circumference at breast height (Cbh) at 1.30 m of soil > 15 cm and total height (Ht) above 1.0 m, were estimated with the aid of a 6-degree ruler, were measured in each plot m in length. Cbh was measured using a tape measure and the data obtained were converted to diameter at breast height (Dbh).

The names of the families were updated according to the proposed Angiosperm Phylogeny Group IV (APG, 2016) system and the scientific names of the species were confirmed according to the nomenclature of the Missouri Botanical Garden (2018) and in "The plant list". The abbreviation of the authors' names followed the one given by Brummitt and Powell (1992). When it was not possible to identify them in the field, the characteristics of the plants and the collections of the botanical material were made for the preparation of exsiccates, with a view to later identification.

Phytosociological Characterization

The structural characterization of the vegetation was performed by means of the analysis of the phytosociological parameters of absolute and relative density (DAi and DRi), absolute and relative frequency (FAi and FRi), absolute and relative dominance (DoAi and DoRi) described by Mueller-Dombois and Ellenberg (1974). Floristic diversity was analyzed using the diversity index of Shannon (H') (Broewer and Zar, 1984) and Pielou equation coefficient (J) (Pielou, 1975).

The analysis of the diametric distribution of individuals in diameter classes was performed by means of histograms, with the number and amplitude between the classes defined by the method of Sturges Method (1926):

$$k = 1 + 3,3 (\log_{10} n); w = R/k$$

where:

k = number of classes;

n = number of observations;

w = total amplitude;

R = difference between the highest and the lowest observed value.

Spatial distribution

The spatial distribution pattern of the species was analyzed for the ten of the highest absolute density observed in the phytosociological analysis of the fragment. For this, the MacGuinnes index (1934) was used to calculate the aggregation pattern of the species by means of the following expressions:

$$IGA_i = \frac{D_i}{d_i}; D_i = \frac{n_i}{UT}; d_i = -\ln\left(1 - \frac{U_i}{UT}\right)$$

Where:

IGA_i = MacGuinnes index;

D_i = observed density of the *i*th species;

d_i = expected density of the *i*th species;

ln = neperian logarithm;

n_i = number of individuals sampled from the *i*th species;

U_i = number of sample units that the *i*th species occurs;

UT = total number of sample units.

Considering that the index is based on the observed and expected density of the species, where IGA_i ≤ 1.0 indicates uniform distribution; IGA_i = 1.0 indicates pattern of random distribution; 1.0 < IGA_i ≤ 2.0 indicates grouping tendency; and IGA_i > 2.0 indicates the grouping of the individuals of the species, considering the ten species with the highest density for analysis purposes, since these represent 63.26% of the number of individuals sampled in the survey.

Statistical analysis

The graphical representation of the histograms of the individuals and data processing were used Microsoft Excel 2010 software and the statistical program R (R Core Team, 2018).

3 RESULTS

Floristic Composition

In the floristic and structural survey of the vegetation, 411 arboreal individuals were cataloged in 58 species belonging to 33 botanical families (Table 1). The most representative families as the richness of species in the studied fragment were Fabaceae with 12 species, followed by Myrtaceae with six, Anacardiaceae with four and Melastomataceae with three species, totaling 43.10% of the species cataloged in the present study.

Table 1: Families and tree species raised in the Atlantic Forest fragment and their respective phytosociological parameters. n_i = number of individuals sampled; DA_i = absolute density in number of individuals (n ha⁻¹); DR_i = relative density (%); FA_i = absolute frequency; FR_i = relative frequency (%); DoA_i = absolute dominance (m² ha⁻¹); DoR_i = relative dominance (%) and VI = value of importance (%).

Family	Botanical name	n _i	DA _i	DR	FA _i	FR	Do _{A_i}	DoR	IV
Chrysobalanaceae	<i>Licania tomentosa</i> (Benth.) Fritsch.	3	12	0.73	10.00	0.68	0.02	0.20	0.54
Hypericaceae	<i>Vismia guianensis</i> (Aubl.) Pers.	1	4	0.24	10.00	0.68	0.02	0.14	0.35
Lecythidaceae	<i>Eschweilera ovata</i> (Cambess.) Mart. ex Miers	43	172	10.46	80.00	5.41	0.74	6.31	7.39
Paulowniaceae	<i>Paulownia fortunei</i> (Seem.) Hemsl.	1	4	0.24	10.00	0.68	0.02	0.14	0.35
Anacardiaceae	<i>Anacardium occidentale</i> L.	2	8	0.49	20.00	1.35	0.05	0.44	0.76
	<i>Erythroxylum deciduum</i> A.St.-Hil.	29	116	7.06	70.00	4.73	0.71	6.05	5.94
	<i>Tapirira guianensis</i> Aubl.	7	28	1.70	40.00	2.70	1.01	8.61	4.34
	<i>Thyrsodium spruceanum</i> Benth.	25	100	6.08	40.00	2.70	0.32	2.68	3.82
Annonaceae	<i>Xylopia frutescens</i> Aubl.	2	8	0.49	20.00	1.35	0.07	0.57	0.80

Apocynaceae	<i>Himatanthus bracteatus</i> (A.DC.) Woodson	4	16	0.97	10.00	0.68	0.14	1.21	0.95
	<i>Plumeria bracteata</i> A.DC.	1	4	0.24	10.00	0.68	0.06	0.55	0.49
Araliaceae	<i>Didymopanax morototoni</i> (Aubl.) Decne. & Planch.	10	40	2.43	60.00	4.05	0.40	3.41	3.30
Boraginaceae	<i>Cordia superba</i> Cham.	1	4	0.24	10.00	0.68	0.01	0.09	0.34
Burseraceae	<i>Protium heptaphyllum</i> (Aubl.) Marchand	3	12	0.73	20.00	1.35	0.07	0.59	0.89
Celastraceae	<i>Maytenus distichophylla</i> Mart.	4	16	0.97	20.00	1.35	0.06	0.55	0.96
Clusiaceae	<i>Clusia nemorosa</i> G.Mey.	19	76	4.62	30.00	2.03	0.82	6.99	4.55
Elaeocarpaceae	<i>Sloanea guianensis</i> (Aubl.) Benth.	1	4	0.24	10.00	0.68	0.02	0.20	0.37
Erythroxylaceae	<i>Erythroxylum citrifolium</i> A.St.-Hil.	4	16	0.97	30.00	2.03	0.05	0.39	1.13
	<i>Erythroxylum revolutum</i> Mart.	1	4	0.24	10.00	0.68	0.02	0.18	0.37
Fabaceae	<i>Albizia Pedicellaris</i> (DC.) L.Rico	1	4	0.24	10.00	0.68	0.30	2.57	1.16
	<i>Abarema filamentosa</i> (Benth.) Pittier	5	20	1.22	30.00	2.03	0.07	0.63	1.29
	<i>Bowdichia virgilioides</i> Kunth	1	4	0.24	10.00	0.68	0.04	0.33	0.42
	<i>Caesalpinia peltophoroides</i> Benthham	48	192	11.68	40.00	2.70	0.83	7.06	7.15
	<i>Cassia Multijuga</i> Rich.	4	16	0.97	20.00	1.35	0.08	0.72	1.01
	<i>Chamaecrista ensiformis</i> (Vell.) H.S.Irwin & Barneby	17	68	4.14	20.00	1.35	0.56	4.73	3.40
	<i>Ormosia bahiensis</i> Monach.	1	4	0.24	10.00	0.68	0.09	0.79	0.57
	<i>Plathymenia foliolosa</i> Benth.	6	24	1.46	40.00	2.70	0.17	1.48	1.88
	<i>Pterocarpus violaceus</i> Vogel	3	12	0.73	20.00	1.35	0.33	2.84	1.64
	<i>Pterodon emarginatus</i> Vogel	7	28	1.70	40.00	2.70	0.26	2.21	2.20
	<i>Sclerolobium paniculatum</i> Vogel	1	4	0.24	10.00	0.68	0.20	1.71	0.88
	<i>Stryphnodendron adstringens</i> (Mart.) Coville	1	4	0.24	10.00	0.68	0.01	0.09	0.34
Goupiaceae	<i>Goupia glabra</i> Aubl.	4	16	0.97	10.00	0.68	0.61	5.23	2.29
Lauraceae	<i>Ocotea glomerata</i> (Nees) Mez	2	8	0.49	20.00	1.35	0.05	0.40	0.75
Lecythidaceae	<i>Lecythis pisonis</i> Cambess.	7	28	1.70	30.00	2.03	0.14	1.19	1.64
Malpighiaceae	<i>Byrsonima crassifolia</i> (L.) Kunth	24	96	5.84	90.00	6.08	0.59	4.98	5.63
Malvaceae	<i>Luehea paniculata</i> Mart.	9	36	2.19	50.00	3.38	0.37	3.13	2.90
Melastomataceae	<i>Miconia hypoleuca</i> (Benth.) Triana	1	4	0.24	10.00	0.68	0.01	0.08	0.33
	<i>Miconia minutiflora</i> (Bonpl.) DC.	1	4	0.24	10.00	0.68	0.01	0.06	0.33
	<i>Miconia prasina</i> (Sw.) DC.	13	52	3.16	70.00	4.73	0.22	1.84	3.24
Moraceae	<i>Brosimum guianense</i> (Aubl.) Huber ex Ducke	6	24	1.46	30.00	2.03	0.06	0.52	1.34
Myrtaceae	<i>Campomanesia dichotoma</i> (O.Berg) Mattos	5	20	1.22	20.00	1.35	0.14	1.20	1.25
	<i>Eugenia</i> sp.	1	4	0.24	10.00	0.68	0.01	0.09	0.34
	<i>Myrcia guianensis</i> (Aubl.) DC.	4	16	0.97	20.00	1.35	0.02	0.21	0.85
	<i>Myrcia sylvatica</i> (G.Mey.) DC.	30	120	7.30	50.00	3.38	0.68	5.80	5.49
	<i>Myrcia tomentosa</i> (Aubl.) DC.	2	8	0.49	20.00	1.35	0.03	0.25	0.70
	<i>Myrcia splendens</i> (Sw.) DC.	12	48	2.92	60.00	4.05	0.34	2.92	3.30
Nyctaginaceae	<i>Bougainvillea spectabilis</i> Willd.	1	4	0.24	10.00	0.68	0.04	0.37	0.43
	<i>Guapira laxa</i> (Netto) Furlan	7	28	1.70	20.00	1.35	0.15	1.24	1.43
Ochnaceae	<i>Ouratea polygyna</i> Engl.	2	8	0.49	10.00	0.68	0.07	0.59	0.58
Olacaceae	<i>Ximenia americana</i> L.	1	4	0.24	10.00	0.68	0.02	0.13	0.35
Peraceae	<i>Pera ferruginea</i> (Schott) Müll.Arg.	8	32	1.95	50.00	3.38	0.33	2.83	2.72
	<i>Pogonophora schomburgkiana</i> Miers ex Benth.	1	4	0.24	10.00	0.68	0.01	0.10	0.34
Polygonaceae	<i>Coccoloba mollis</i> Casar.	4	16	0.97	30.00	2.03	0.08	0.65	1.22
Primulaceae	<i>Myrsine guianensis</i> (Aubl.) Kuntze	4	16	0.97	20.00	1.35	0.05	0.42	0.91
Rubiaceae	<i>Alseis pickelle</i> Pilg. & Schmale	2	8	0.49	10.00	0.68	0.10	0.81	0.66
Rutaceae	<i>Zanthaxilum</i> sp.	1	4	0.24	10.00	0.68	0.02	0.20	0.37
Salicaceae	<i>Casearia javitensis</i> Kunth	2	8	0.49	20.00	1.35	0.03	0.24	0.69
Sapindaceae	<i>Allophylus edulis</i> (A.St.-Hil., A.Juss. & Cambess.) Radlk.	1	4	0.24	10.00	0.68	0.01	0.07	0.33
Total		41	164	100.0	1480.	100.0	11.7	100.0	100.0
		1	4	0	00	0	5	0	0

We sampled 1644 individuals, where the most representative species for absolute and relative density were *Caesalpinia peltophoroides*, *Eschweilera ovata*, *Myrcia sylvatica*, *Erythroxylum deciduum*, *Thyrsodium spruceanum* and *Byrsonima crassifolia*, totaling 48.42% of the individuals sampled (Figure 3).

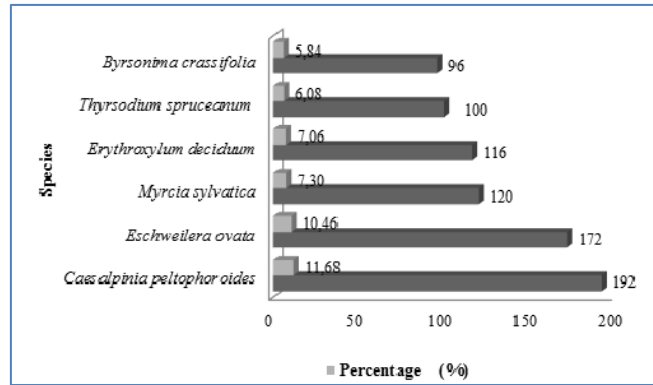


Fig. 3: Species that presented higher values of DAI and DRi in the Atlantic Forest fragment.

The species that presented the highest value of importance were *Eschweilera ovata*, *Erythroxylum deciduum*, *Byrsonima crassifolia*, *Myrcia sylvatica* and *Clusia nemorosa* (Figure 4).

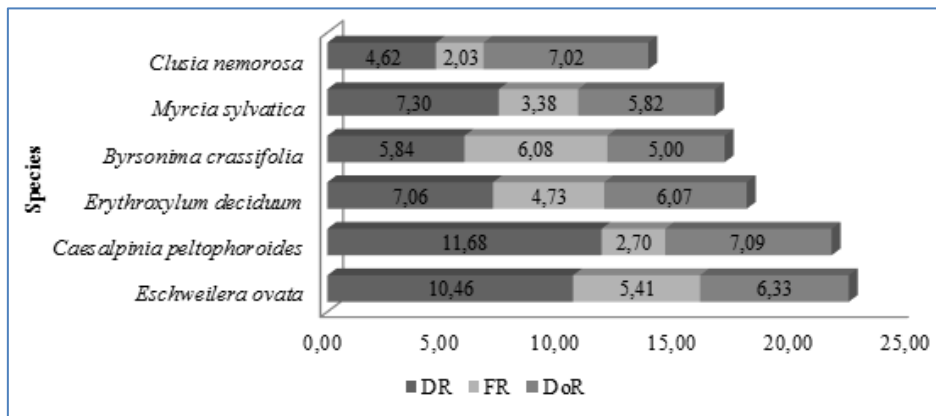


Fig. 4: Relation of species with higher Value of Importance (VI) in the fragment.

The arithmetic mean of the diameters of the arboreal individuals sampled in the fragment was 7.78 cm, with a diametric variation between 3.66 and 34.70 cm. *Goupia glabra* was the species with the highest DBH (34.70 cm), followed by *Tapirira guianensis* (34.54 cm), *Sclerolobium paniculatum* (25.31 cm) and *Clusia nemorosa* (24.83 cm). The remainder presented a diametric distribution pattern tending to the J-inverted, represented by a higher frequency of individuals concentrated in smaller diameter classes and the progressive decrease of the frequency with the increase of the diametric class (Figure 5).

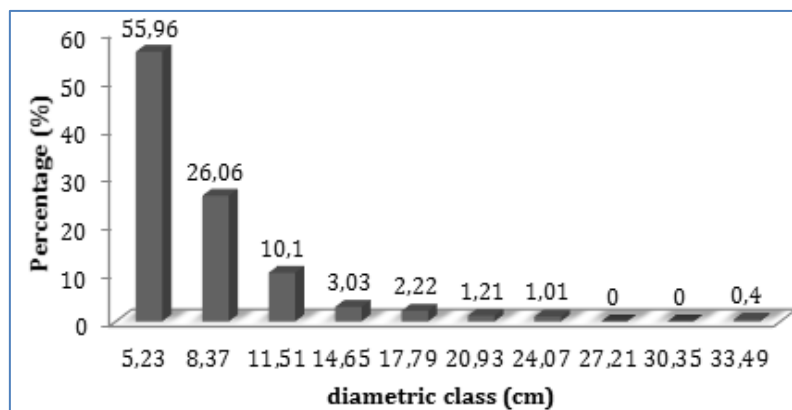


Fig. 5: Diameter distribution of tree individuals with DBH ≥ 4.77 cm (Cbh ≥ 15 cm) sampled.

The Shannon diversity index for the forest community was 3.37 nats. ind.⁻¹ and Pielou equability index was 0.83. In general, the species presented an aggregate pattern of distribution in the area, with emphasis on the species *Miconia prasina* and *Myrcia splendens*, which presented a pattern with tendency to cluster and *Byrsonima crassifolia* that presented a random pattern (Table 2).

Table 2. Pattern of spatial distribution of species in the forest fragment.

Species	IGA	Classification
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<i>Caesalpinia peltophoroides</i>	9.4	Aggregate distribution
<i>Eschweilera ovata</i>	2.7	Aggregate distribution
<i>Myrcia sylvatica</i>	4.3	Aggregate distribution
<i>Erythroxylum deciduum</i>	2.4	Aggregate distribution
<i>Thyrsoodium spruceanum</i>	4.9	Aggregate distribution
<i>Byrsonima crassifolia</i>	1.0	Random distribution
<i>Clusia nemorosa</i>	5.3	Aggregate distribution
<i>Chamaecrista ensiformis</i>	7.6	Aggregate distribution
<i>Miconia prasina</i>	1.1	Tendency to cluster
<i>Myrcia splendens</i>	1.3	Tendency to cluster

4 DISCUSSION

As in the present study, other authors such as Araújo *et al.* (2015), Colonetti *et al.* (2009) and Costa Junior *et al.* (2008) analyzing the phytosociological structure of fragments of Atlantic Forest also showed that a small number of species presented high density. According to Richards (1996), the dominance of a reduced number of species in the environment is a common occurrence in humid tropical forests. However, this fact may also be related to the successional stage of the fragment, where few species dominate the environment and the resources of the environment. According to Martins (2012), with vegetation advancement in the process of ecological succession, the density of individuals tends to decrease due to greater inter and intraspecific competition.

Sawczuk *et al.* (2014), analyzing the changes in the horizontal structure of a mixed ombrophilous forest in the state of Paraná, showed that the progression of the successional stage was accompanied by a decrease in the total number of trees, whose authors associated such response to factors such as inter and intraspecific and to a greater dominance exerted by the individuals of *Araucaria angustifolia* and *Ocotea porosa*, that in turn, increased their diameters in the period of study between the years of 2002 and 2008. However, to test this hypothesis, studies related to the population dynamics of tree species are suggested in order to follow the changes in the structure of the remnant.

The absolute dominance in the community was 11.75 m² ha⁻¹, being lower than that found by Costa Junior *et al.* (2008), Araújo *et al.* (2015) and Lima *et al.* (2017) when analyzing the structure of Atlantic Forest fragments in the Brazilian Northeast, whose values were 22.13 m² ha⁻¹, 23.59 m² ha⁻¹ and 19.63 m² ha⁻¹, respectively. This fact can be explained by the successional stage of the fragment of forest analyzed, the size of the area sampled, as well as the anthropic disturbances that preceded the incorporation of the fragment to the Conservation Unit Dois Irmãos State Park.

According to Martins (2012) dominance is an attribute represented by abundance expressed in m² ha⁻¹, having been used worldwide to represent vegetation biomass. In order to analyze the species' dominance of the community, it was observed that *Tapirira guianensis*, *Caesalpinia peltophoroides*, *Clusia nemorosa*, *Eschweilera ovata*, *Erythroxylum deciduum* and *Myrcia sylvatica* are the species that contribute the most to the biomass of the studied fragment, totaling 40.96%.

The species that obtained the highest frequency in the survey were *Byrsonima crassifolia*, *Eschweilera ovata*, *Miconia prasina* and *Erythroxylum deciduum*. According to Araújo *et al.* (2015) a relatively uniform distribution of species in vegetation may be related to dispersal strategies, species spatial pattern, species adaptation to limiting factors, ecological behavior and inter- and intraspecific interactions among the species of the community.

The species that most stood out in Valor of Importance was *Eschweilera ovata*, by the highest values of relative density and relative frequency, being present in 80% of the plots sampled. *Caesalpinia peltophoroides* was present in 40% of the plots, resulting in the second highest value of Relative Dominance (Table 1). Oliveira *et al.* (2013) evaluated the horizontal structure in a fragment of Atlantic Forest in the State of Pernambuco and also identified the species *Eschweilera ovata* as one of the species with the highest Importance Value. Pinto *et al.* (2018) in Igarassu, Pernambuco, evaluated the structure of a secondary forest fragment in the São José Plant and found lower values of VI, indicating a late succession, corroborating with the good state of conservation of the fragment evaluated in the present study.

In relation to the diameter distribution of the individuals in the study area, other works carried out on the same typology, as that of Araújo *et al.* (2015), Marchiori *et al.* (2016), Cysneiros *et al.* (2017), Lima *et al.* (2017) also showed a higher proportion of individuals in smaller diameter classes, indicating that it is a factor associated with the natural dynamics in multiannual forests, whose high number of regenerating individuals compensates for mortality over time (Dalla Lana *et al.* 2013).

As for the floristic diversity of the area, the results are similar to other fragments of the Atlantic Forest, such as 3.19 nats. ind.⁻¹ (Araújo *et al.*, 2015); 2.9 nats. ind.⁻¹ (Pinto *et al.*, 2018); and 3.19 nats. ind.⁻¹ (Cordeiro *et al.*, 2017). However, it is lower than those found by Campos *et al.* (2011), Lima *et al.* (2017) and Marchiori *et al.*, (2016), who verified values of 4.05 nats. ind.⁻¹; 4.0 nats. ind.⁻¹ and 3.7 nats. ind.⁻¹, respectively. This lower diversity when compared to the aforementioned works may be related to the initial successional stage of the remnant, where the structural complexity is smaller when compared to more advanced fragments in the process of ecological succession.

In relation to the species distribution patterns in the area, tropical forests generally present high spatial and temporal heterogeneity, which strongly influences species distribution patterns (Gris *et al.*, 2014). In tropical ombrophilous forests, restricted seed dispersal around the maternal plant may be a preponderant factor for the spatial pattern found for species with aggregate distribution (Fajardo *et al.*, 2015). Varella *et al.* (2018) emphasize that understanding the structure and spatial distribution of tree species is necessary to predict the spatial variation of successional, genetic and ecological processes of a forest

ecosystem. According to Ferreira et al. (2016), the behavior of the spatial distribution reflects the autochoric dispersion syndrome, which confers the aggregate behavior in early-stage populations and a random pattern in the more advanced stage of development.

5 CONCLUSIONS

The phytosociological analysis of the vegetation allowed to know the floristic composition and characterization of the horizontal structure of the remnant, where despite a diametric distribution tending to the inverted J, indicating that the forest is in the process of regeneration, a small number of species stand out over while others are scarce in the remnant, deserving prominence in the conservation of biological diversity, so that actions of inadequate management do not imply extinction.

With the exception of *Byrsonima crassifolia* that presented a random distribution in the fragment and the species *Miconia prasina* and *Myrcia splendens*, the other species whose aggregation pattern was analyzed presented an aggregated spatial distribution of the individuals. This fact allows to understand the behavior of these species, to subsidize projects that use them, aiming to recover degraded areas in the remainder studied, as well as to establish strategies for the conservation of floristic diversity.

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